CO₂ fertilization in temperate FACE experiments not representative of boreal and tropical forests

THOMAS HICKLER*, BENJAMIN SMITH*, I. COLIN PRENTICE[†], KRISTINA MJÖFORS^{*1}, PAUL MILLER*, ALMUT ARNETH* and MARTIN T. SYKES*

*Department of Physical Geography and Ecosystems Analysis, Geobiosphere Science Centre, Lund University, Sölvegatan 12, S-223 62 Lund, Sweden, †Department of Earth Sciences, University of Bristol, Wills Memorial Building, Queen's Road, Clifton, Bristol BS8 1RJ, UK

Abstract

Results from free-air CO₂ enrichment (FACE) experiments in temperate climates indicate that the response of forest net primary productivity (NPP) to elevated CO₂ might be highly conserved across a broad range of productivities. In this study, we show that the LPJ-GUESS dynamic vegetation model reproduces the magnitude of the NPP enhancement at temperate forest FACE experiments. A global application of the model suggests that the response found in the experiments might also be representative of the average response of forests globally. However, the predicted NPP enhancement in tropical forests is more than twice as high as in boreal forests, suggesting that currently available FACE results are not applicable to these ecosystems. The modeled geographic pattern is to a large extent driven by the temperature dependence of the relative affinities of the primary assimilation enzyme (Rubisco) for CO₂ and O₂.

Keywords: carbon cycle, climate change, CO₂ fertilization, ecosystem model, FACE, forest response, LPJ-GUESS, NPP, photosynthesis, Rubisco

Received 13 December 2006; revised version received 15 August 2007 and accepted 13 September 2007

Introduction

Increasing atmospheric CO₂ is one of the most important global change drivers, not only as an agent of greenhouse forcing but also as a limiting resource for carbon assimilation by plants (Farquhar et al., 1980). Numerous small-scale greenhouse or open-top chamber experiments have shown that an increase in CO₂ concentrations of about 300 ppmv enhances photosynthesis in C_3 plants on the order of 60%, leading to substantial, although more variable, increases in net primary productivity (NPP; for reviews, see Curtis & Wang, 1998; Norby et al., 1999). However, the applicability of such results to real ecosystems has been questioned (Norby et al., 1999). Free-air CO₂ enrichment (FACE) technology was developed in order to overcome the recognized problems of smaller scale experiments (Hendrey et al., 1999; Norby et al., 1999). FACE results generally confirm

Correspondence: Thomas Hickler, e-mail: Thomas.hickler@nateko.lu.se

¹Present address: Kristina Mjöfors, Department of Forest Soils, Swedish University of Agricultural Science, PO Box 7001, S-750 07 Uppsala, Sweden. the enhancement of NPP by elevated CO₂ (Nowak *et al.*, 2004; Ainsworth & Long, 2005; Norby *et al.*, 2005). After analyzing the results from the longest running forest FACE experiments, Norby *et al.* (2005) concluded that '... the response of forest NPP is highly conserved across a wide range of productivity, with a stimulation at the median of $23 \pm 2\%'$ (for approximately 550 ppmv CO₂).

Studies with large-scale ecosystem models, which simulate physiological CO2 effects based on the biochemical processes underlying carbon assimilation, have suggested that CO₂ 'fertilization' could be having a substantial impact on global NPP (Melillo et al., 1993; Cramer et al., 2001). The positive effects of increasing CO₂ might also be a major driver of the land carbon sink during the 1980s and 1990s (Prentice et al., 2001). The magnitude of these modeled CO₂ effects has been contested; however (e.g. Hungate et al., 2003), especially as models tend to predict a still larger NPP enhancement with a continuing CO_2 rise into the future (Pan et al., 1998; Cramer et al., 2001; McGuire et al., 2001; Bachelet et al., 2003). A necessary test for models is therefore to evaluate the magnitude of their response to the same CO_2 enhancement as applied in FACE experiments.

In this study, we compared the response of forest NPP that is predicted by the dynamic vegetation model LPJ-GUESS (Smith *et al.*, 2001) to the results from FACE experiments, as summarized by Norby *et al.* (2005). We also carried out a global forest FACE experiment and analyzed geographic differences in the predicted CO_2 response of forests, including climate regions for which no experimental data are available. We restricted our study to forest ecosystems, as these have the largest potential of influencing the global carbon cycle (Bousquet *et al.*, 2000; Prentice *et al.*, 2001).

Material and methods

LPJ-GUESS

LPJ-GUESS (Smith et al., 2001; Hickler et al., 2004) is a generalized, process-based model of vegetation dynamics and biogeochemistry designed for regional to global applications. It combines features of the widely used Lund-Potsdam-Jena Dynamic Global Vegetation Model (LPJ-DGVM; Sitch et al., 2003) with those of the General Ecosystem Simulator (GUESS; Smith et al., 2001) in a single, flexible modeling framework. The models have identical representations of ecophysiological and biogeochemical processes, including the hydrological cycle updates described in Gerten et al. (2004) and disturbance by wildfires (Thonicke et al., 2001). They differ in the level of detail with which vegetation dynamics and canopy structure are simulated: simplified but computationally efficient representations are used in the LPJ-DGVM, while in GUESS, a more detailed and mechanistic approach, distinguishing individual and population structure and patch-scale heterogeneity, is used. Model representations of stochastic establishment, individual tree mortality, and disturbance events in GUESS are based on the FORSKA2 forest gap model (Prentice et al., 1993). In order to represent regional vegetation, a number of replicate patches (here 50) are simulated, and the average is taken to represent the vegetation at a stand to landscape scale.

Physiological processes (e.g. photosynthesis, plant respiration, and microbial decomposition) and associated fluxes of carbon and water between soil layers, vegetation, and the atmosphere in LPJ-GUESS are simulated on a daily time step. Growth and vegetation dynamics are updated by allocating the annually accrued NPP to leaves, sapwood, and fine roots in conformity with a set of allometric rules (Sitch *et al.*, 2003). Sapwood-to-heartwood conversion, litterfall, fine root turnover, establishment, mortality, and disturbance are all simulated annually. Vegetation is represented as a mixture of plant functional types (PFTs), which are differentiated by bioclimatic limits and physiological, morphological, phenological, and life history criteria governing competition for light and water. The model is driven by daily values of temperature, precipitation, and percentage sunshine hours, information on latitude, and soil texture, and a global atmospheric CO_2 concentration value.

A modified Farquhar photosynthesis scheme (Collatz et al., 1991; Haxeltine & Prentice, 1996a, b) calculates vertically integrated canopy photosynthesis analytically as a function of absorbed photosynthetically active radiation, temperature, and atmospheric CO₂ concentration, under the assumption of optimal nitrogen allocation to leaves at different levels in the canopy (Haxeltine & Prentice, 1996a). If the water supply, which is determined by plant root-weighted soil moisture availability and maximum sapflow rates, is lower than the atmospheric demand, which is calculated based upon the leaf conductance associated with an unstressed photosynthesis rate, canopy conductance is reduced until transpiration equals the supply. Under these circumstances, the diffusion of CO₂ into the leaf is also decreased, resulting in lower photosynthesis rates.

In this study, a generic patch-destroying disturbance regime, representing stochastic events such as windstorms and pest outbreaks, was applied with a mean disturbance interval of 100 years. The set of PFTs in the LPJ-DGVM was enhanced by subdividing each PFT into shade-tolerance classes (Smith *et al.*, 2001; Hickler *et al.*, 2004). Vegetation dynamics parameters were re-calibrated using data on pristine forest vegetation dynamics, composition, and structure in different biomes (P. Miller & T. Hickler, unpublished results).

LPJ-GUESS and the closely related LPJ-DGVM have been evaluated by comparison with observations in a large number of studies (http://www.pik-potsdam.de/ lpj). Output variables evaluated include NPP (e.g. Zaehle *et al.*, 2005; Hickler *et al.*, 2006), stand-scale and continental-scale evapotranspiration (AET) and runoff (Gerten *et al.*, 2004), vegetation greening trends in high northern latitudes (Lucht *et al.*, 2002) and the African Sahel (Hickler *et al.*, 2005), stand-scale leaf area index (LAI) and gross primary productivity (GPP; Arneth *et al.*, 2007), and forest stand structure and development (Smith *et al.*, 2001; Hickler *et al.*, 2004).

Model experiments

We evaluated the model through a site-by-site comparison with FACE experiments, and by carrying out a 'global forest FACE experiment' with the model.

For four FACE experiments [the same as in Norby *et al.* (2005); Table A1], the model was run with site-specific meteorological data (temperature, precipitation,

and incoming shortwave radiation), as well as information on soil texture, CO₂ concentration in the atmosphere, and vegetation structure and composition. Meteorological data were obtained through site home pages [FACTS II (http://www.fs.fed.us/nc/face (date of access: April 2004)), ORNL (http://public.ornl.gov/ face/ORNL (date of access: April 2004))] or directly from researchers working at the sites. The climate input consisted of daily fields, except for FACTS I, where daily values were derived by linear interpolation between monthly average values. As site-specific meteorological data were only available during the course of the CO₂ experiments, a gridded global climate dataset (CRU TS 2.1; Mitchell & Jones, 2005), was used for the period of forest growth before CO₂ elevation began. For this period, daily values were derived by linear interpolation between the monthly means.

The choice of years for the data-model intercomparison (Fig. 1) was constrained by the availability of meteorological data at the time of the modeling experiments and published NPP estimates. The years used in this study are therefore not identical to those included by Norby *et al.* (2005), who only used data from the years after canopy closure. Using this criterion, Norby *et al.* (2005) excluded the years that we used at FACTS II, and the first 2 years of data that we used for ORNL.



Fig. 1 Simulated vs. observed annual NPP enhancement (mean and 95% confidence intervals) under elevated CO_2 for four FACE sites, averaged over the period for which station climate and NPP data were available: 1999–2002 for ORNL (Oak Ridge National Laboratory), temperate broad-leaved forest (observed NPP from Norby *et al.*, 2002, 2003); 1998–2000 for FACTS I (forest atmosphere carbon transfer and storage I), North Carolina, temperate needle-leaved forest (Schäfer *et al.*, 2003); 1999–2001 for FACTS II, Wisconsin, boreal broad-leaved forest (King *et al.*, 2005); 2000–2001 for PopFACE, Italy, broad-leaved coppice plantation (Gielen *et al.*, 2005). At the PopFACE site, data from different species (Table A1) were used as independent data points for calculating the confidence interval (N = 6).

Modeled and observed vegetation structure and composition are given in Table A1. The growing conditions at the FACE sites (monocultures with a certain planting density and stand age) were prescribed in the model (Table A1), and further establishment was disabled up to and during the period of CO_2 enhancement. In the case of the two older stands (FACTS 1 and ORNL), the model was run with the standard model representations of natural tree mortality. Mortality was switched off for the FACTS II and PopFACE site, because the stands are only a few years old and substantial mortality has not occurred. LAI was allowed to respond to CO_2 , even if the experimental data do not suggest LAI changes.

The tree species planted at the FACE sites are not explicitly parameterized in the model and were represented by the corresponding PFT (Table A2). The main differences between the PFTs at the four different sites concern phenology (evergreen vs. deciduous), leaf or needle form (e.g. specific leaf area), and climatic distribution (boreal vs. temperate), whereby boreal types have a lower optimum temperature for photosynthesis (Smith *et al.*, 2001) and higher base respiration rates. More information on PFT parameterization is given in Table A3.

At FACTS II, model results were compared with the observed response of pure trembling aspen (*Populus tremuloides*) stands, because differences in the strength of intra-specific and inter-specific competition might influence the observed CO_2 response (King *et al.*, 2005).

At the PopFACE site, irrigation (Calfapietra *et al.*, 2003) was implemented within the model by increasing daily precipitation until it matched the modeled daily potential evapotranspiration (PET; Haxeltine & Prentice, 1996b), if PET was larger than rainfall during the growing season (mean daily temperature above 5 °C). Model results were only compared with the results from the first rotation period (Calfapietra *et al.*, 2003). Plant allometric relationships and thereby growing conditions during the second rotation, when a developed root system was in place from the first rotation period (Liberloo *et al.*, 2006), are not represented in the model in its current form.

The generalized parameterization of the model was not adapted to site conditions, with the exception of the leaf area to sapwood cross-sectional area ratio of the PFT representing *Pinus taeda* at FACTS I (Table A3). Without more site-specific and species-specific calibration, which is beyond the scope of this study, we would not expect the model to reproduce details in the observed variation of CO_2 responses. However, the results of the global FACE experiment can only be regarded realistic if the model reproduces the general magnitude of the NPP response to elevated CO_2 .

The global model experiment was carried out using the gridded CRU TS 2.1 global climate dataset (Mitchell & Jones, 2005), and gridded information on soil texture (Sitch et al., 2003). Two runs were carried out: one with actual historical CO₂ concentrations from 1901 to 2002 [McGuire et al., 2001; TRENDS (http://cdiac.esd.ornl. gov/trends/co2/contents.htm)], and one with historical CO₂ concentrations increased to 550 ppmv during 1996-2002. The latter 6-year period was chosen for analysis of the model results, as it would represent climatic conditions similar to those prevailing during the period when most of the FACE data were obtained. The model was run with potential natural vegetation and for all grid cells that would naturally carry forests (Haxeltine & Prentice, 1996b), with the exception of tropical deciduous forests, which have a savanna-like structure. Following a 300-year spin-up to establish the 'steady-state' vegetation, the model was driven by the observed climate from 1901 to 2002.

Results

Modeled and observed NPP responses to elevated CO₂ may be compared in Fig. 1. Averaged over all sites, NPP was increased by 27.8% ($\pm 10.0\%$; N = 4) according to the FACE data, and the model predicted a 25.9% ($\pm 6.8\%$; N = 4) increase in NPP. However, at FACTS I, the modeled response was 10.3% higher than that observed in absolute terms (corresponding with an overestimation by 49%), while at FACTS II, the modeled response was 24.6% lower than the observed one (Fig. 1).

Table 1 compares the CO_2 response of forest NPP to elevated CO_2 indicated by Norby *et al.* (2005) to the response predicted by the model, averaged over all areas of the globe that are potentially forested, and for different biomes. The modeled global average response

Table 1Modeled NPP response (averages over 1996–2002and 95% confidence intervals) as predicted for all forestsglobally and for different biomes, compared with the responseof forest NPP suggested by Norby *et al.* (2005)

| | NPP enhancement (%) | Ν |
|---------------------|---------------------|---------|
| Norby et al. (2005) | 23 ± 2 | 19 |
| Global forests | 24.5 ± 0.06 | 130 571 |
| Boreal forests | 15.1 ± 0.06 | 77 159 |
| Temperate forests | 25.7 ± 0.14 | 22 974 |
| Tropical forests | 35.1 ± 0.09 | 30 4 39 |

For calculating confidence intervals (two times the standard error of the sample mean), different years of data or modeled response for the same site or grid cell were treated as independent data points. was within the confidence interval of the estimate by Norby *et al.* (2005), based on four FACE forests. The modeled average response for temperate forests was 2.7% higher than that suggested by Norby *et al.* (2005). More strikingly, the modeled response in tropical forests was more than twice as high as the response predicted for boreal forests.

Figure 2 shows the geographical pattern of the modeled NPP enhancement in response to elevated CO₂. The NPP enhancement was predicted to generally increase toward the equator. The strongest response was predicted for warm temperate and tropical forests. Geographic variations in the NPP response [Δ NPP (%)] were largely associated with differences in temperature, linear regression analysis yielding $R^2 = 0.67$ (Fig. A1a). Water availability, expressed as the ratio of actual to potential evapotranspiration (AET/PET) for ambient CO₂ concentrations, was also significantly related to Δ NPP, but only with $R^2 = 0.11$ (Fig. A1b). The residuals from the regression with temperature were only weakly related to AET/PET ($R^2 = 0.01$), with a stronger NPP response under drier conditions. Because of the large sample size (21762 global forest grid cells), all regressions are highly significant (P < 0.001), but model results based on interpolated gridded spatial climate datasets as here are spatially highly auto-correlated, making the P-value meaningless.

Grid cells for which the model predicted very low NPP ($0.1 \text{ kg m}^{-2} \text{ yr}^{-1}$) and grid cells for which the model did not predict forest biomes (woody LAI < 0.5 for boreal forests and 2.5 for the remaining forests; Haxeltine & Prentice, 1996b) were excluded. These restrictions had only a very minor effect on the presented results.

Discussion

The model reproduced the overall response of forest productivity to elevated CO2 observed in FACE experiments. We regard this as a requirement for any ecosystem model that is to be applied for simulating future ecosystem dynamics under rising levels of CO₂. Part of the discrepancies between the modeled and observed response at individual sites might be explained by the fact that the model parameterization was not explicitly adapted for all site-specific conditions that might influence the response to elevated CO₂; for example, the actual planted tree species. The number of sites, however, is too small to draw conclusions as to the general ability of the model to capture between-site variations in response to CO₂. In the case of FACTS I, the model might have overestimated the response, because nutrient limitations, which are not explicitly represented in the model, might limit the NPP enhancement under



Fig. 2 Geographic pattern of the simulated NPP enhancement, resulting from a step increase of CO₂ from ambient to 550 ppmv; NPP values averaged over 1996–2002.

elevated CO_2 (Oren *et al.*, 2001; Finzi *et al.*, 2002, 2006; McCarthy *et al.*, 2006). At the FACTS II site, the reasons for the mismatch are less clear. King *et al.* (2005) speculated that the response at this site might be particularly strong, because the recently weathered soils may be high in nutrients, compared with many temperate forests, and because competition for light is less intense in the relatively open canopy. The observed NPP enhancement increased from 29% in 1999 to 64% in 2001 (King *et al.*, 2005; Table 3), while the modeled response was rather constant over the same period. This indicates that the model might not capture the growth conditions in this young forest plantation, possibly because LAI and NPP are overestimated (Table A1).

Geographic variations

The model predicted that CO_2 effects on NPP differ between biomes (Table 1; Fig. 2). Climate change is likely to increase NPP more in cold northern regions than close to the equator because of a greater proportional growing season extension in temperature-limited environments (Rustad *et al.*, 2001; Morales *et al.*, 2007). The results of this study suggest that the direct CO_2 response of NPP, by contrast, will be stronger in warm regions.

Regional differences in the modeled CO₂ response were to a large extent driven by the temperature response of the relative affinity of the carboxylation enzyme Rubisco for CO₂ and O₂. Increased temperatures lead to an increased relative fixation of O2 (oxygenation), leading to photorespiration, an energy-dependent process that reduces net photosynthesis (Jordan & Ogren, 1984). Using the models of Farquhar et al. (1980) and Farquhar & von Caemmerer (1982), Long (1991) showed that this effect causes a much stronger CO₂ enhancement of photosynthesis at high temperatures than at low temperatures. According to the model, an increase in atmospheric CO2 from 350 to $650 \,\mu\text{mol}\,\text{mol}^{-1}$ could increase light-saturated CO₂ uptake by 20% at 10 °C and by 105% at 35 °C. The version of the Farquhar model which is implemented within LPJ-GUESS likewise predicts such a strong temperature dependence of the photosynthetic response on elevated CO₂ (Fig. A2). Kirschbaum (1994) coupled a simplified version of the Farquhar model to a model of stomatal

conductance and predicted a similar geographic pattern in response of photosynthesis to elevated CO_2 as found for NPP in the present study. Thus, there exists a strong physiological basis for a temperature dependence of NPP responses on elevated CO_2 .

LPJ-GUESS tends to simulate a stronger response to CO_2 under drier conditions, because elevated CO_2 reduces the negative effects of drought on plant growth (Gerten *et al.*, 2005), which has limited support by the available data (Wullschleger *et al.*, 2002; Nowak *et al.*, 2004). Temperature, however, is a stronger driving force in the model.

Interactions between temperature and CO₂ have recently received considerable attention in the literature, but no quantitative conclusions have been drawn (Norby & Luo, 2004), perhaps because of a paucity of unequivocal data from field experiments (Morison & Lawlor, 1999; Norby & Luo, 2004). In this regard, it may be unfortunate that current long-running forest FACE experiments are all located in the temperate zone and hence subject to a limited range of growing season temperatures. Manipulating air temperature in forests is difficult, and therefore most forest data with multiple temperature treatments come from soil warming experiments or from small numbers of young trees (Norby & Luo, 2004). Furthermore, very few experiments have addressed growth stimulation by CO₂ at temperatures below 10 °C (Morison & Lawlor, 1999), which can prevail during a significant part of the growing season in cold temperate and boreal forests. The predicted weakness of CO2 effects at low temperatures might partly explain why rather small CO₂ effects have been found in high alpine (Körner et al., 1996) and arctic (Tissue & Oechel, 1987) ecosystems.

Long-term responses

It remains to be seen whether the strong medium-term responses observed in FACE experiments will be maintained in the longer term, and if the results obtained in young plantations also apply to mature forests (Körner, 2000). It has been argued that nutrient limitations (especially N) will constrain NPP responses in the future (Finzi et al., 2002; Hungate et al., 2003), but to date there is no unambiguous evidence for a general effect of this kind (Finzi et al., 2006; Norby & Iversen, 2006). In the FACTS I experiment, for example, elevated CO₂ has now increased basal area increment (BAI) by 13–27% for 8 years, with no reduction over time (Moore et al., 2006), even though the response might have been even stronger in the absence of nutrient limitation (Oren et al., 2001). Finzi et al. (2007) showed that at FACTS I, FACTS II, and ORNL, N uptake rates were increased under elevated CO_{2} , which could be one reason for the prolonged NPP enhancement in spite of strong nutrient limitation at FACTS I and ORNL.

Regarding the applicability of FACE results from young plantations to mature forests, Körner et al. (2005) and Asshoff et al. (2006) did not find significant increases in BAI in a mature deciduous forest in Switzerland. However, the photosynthetic response was similar to that of younger stands (Zotz et al., 2005), only a small number of adult trees were measured, and NPP was not estimated. McCarthy et al. (2006) showed that NPP responses at FACTS I were mainly controlled by changes in LAI, with no NPP effect after canopy closure for a given LAI, but further experiments are necessary to establish the generality of this result. One also has to keep in mind that most forests are not mature, being constantly altered by humans or natural disturbances (e.g. Botkin & Simpson, 1990). In summary, current evidence does not preclude the possibility that strong long-term future impacts of increasing CO₂ on NPP and carbon storage, as predicted by the LPJ model in a number of studies (e.g. Cramer et al., 2001; Bachelet et al., 2003), might be realistic.

Final remarks

This study clearly demonstrates the benefits arising from the combination of experimental work and modeling. Process-based models such as LPJ-GUESS are primarily based upon mechanistic understanding of ecosystem processes, such as photosynthesis and autotrophic respiration, which has largely been obtained through experimental work at small scales. Outcomes of modeling studies can be independently tested against results from ecosystem scale experiments and, if the test is successful, can be used to extrapolate beyond the unavoidably limited environmental space of the experiments. Process-based ecosystem models can include the effects of simultaneous variations in multiple environmental drivers, which can rarely be investigated in ecosystem experiments (Norby & Luo, 2004), but they can also be used to distinguish the effects of individual drivers (Lucht et al., 2002; Hickler et al., 2005). In this study, the model has generated a hypothesis about the magnitude of the CO₂ response in different climatic regions, which can be tested by applying FACE methodology in tropical and boreal forests. Factors, which are not explicitly represented by the model, such as nutrient limitations, might lead to a different geographical pattern.

Acknowledgements

Karina Schäfer, Ram Oren, Carlo Calfapietra, Victoria Wittig, and Richard Norby have kindly provided climate data for FACTS I and PopFACE, as well as useful information on the experimental setup at these sites. Timothy D. Mitchell from the Climatic Research Unit, University of East Anglia, provided the CRU 2.1 gridded climate dataset. Thomas Hickler was partly funded by the European Union 6th framework program ENSEMBLES (contract no. 505539). Benjamin Smith acknowledges support from the Swedish Research Council for Agricultural Sciences, Environment and Spatial Planning; and Paul Miller acknowledges funding from the EU FP6 project ALARM (contract no. GOCE-CT-2003-506675).

References

- Ainsworth EA, Long SP (2005) What have we learned from 15 years of free-air CO₂ enrichment (FACE)? A meta-analytic review of the responses of photosynthesis, canopy properties and plant production to rising CO₂. *New Phytologist*, **165**, 351–372.
- Arneth A, Niinemets U, Pressley S et al. (2007) Process-based estimates of terrestrial ecosystem isoprene emissions. Atmospheric Chemistry and Physics Discussions, 7, 31–53.
- Asshoff R, Zotz G, Körner C (2006) Growth and phenology of mature temperate forest trees in elevated CO₂. *Global Change Biology*, **12**, 848–861.
- Bachelet D, Neilson RP, Hickler T *et al.* (2003) Simulating past and future dynamics of natural ecosystems in the United States. *Global Biochemical Cycles*, **17**, 1045, doi: 10.1029/2001GB001508.
- Botkin DB, Simpson LG (1990) Biomass of the North American boreal forest: a step toward accurate global measures. *Biogeochemistry*, 9, 161–174.
- Bousquet P, Peylin P, Ciais P, Le Quere C, Friedlingstein P, Tans PP (2000) Regional changes in carbon dioxide fluxes of land and oceans since 1980. *Science*, **290**, 1342–1346.
- Calfapietra C, Gielen B, Galema ANJ *et al.* (2003) Free-air CO₂ enrichment (FACE) enhances biomass production in a short-rotation poplar plantation. *Tree Physiology*, **23**, 805–814.
- Collatz GJ, Ball JT, Grivet C, Berry JA (1991) Physiological and environmental regulation of stomatal conductance, photosynthesis and transpiration: a model that includes a laminar boundary layer. *Agricultural and Forest Meteorology*, **54**, 107–136.
- Cramer W, Bondeau A, Woodward FI *et al.* (2001) Global response of terrestrial ecosystem structure and function to CO₂ and climate change: results from six dynamic global vegetation models. *Global Change Biology*, **7**, 357–373.
- Curtis PS, Wang XS (1998) A meta-analysis of elevated CO₂ effects on woody plant mass, form, and physiology. *Oecologia*, **113**, 299–313.
- Farquhar GD, von Caemmerer S (1982) Modeling of photosynthetic response to environmental conditions. In: *Physiological Plant Ecology II: Water Relations and Carbon Assimilation* (eds Nobel PS, Osmond CB, Ziegler H), pp. 549–587. Springer, Berlin.
- Farquhar GD, von Caemmerer S, Berry JA (1980) A biochemical model of photosynthetic CO₂ assimilation in leaves of C₃ plants. *Planta*, **149**, 78–90.
- Finzi AC, DeLucia EH, Hamilton JG, Richter DD, Schlesinger WH (2002) The nitrogen budget of a pine forest under free air CO₂ enrichment. *Oecologia*, **132**, 567–578.

- Finzi AC, Moore DJP, DeLucia EH *et al.* (2006) Progressive nitrogen limitation of progressive ecosystem processes under elevated CO₂ in a warm-temperate forest. *Ecology*, **87**, 15–25.
- Finzi AC, Norby RJ, Calfapietra C *et al.* (2007) Increases in nitrogen uptake rather than nitrogen-use efficiency support higher rates of temperate productivity under elevated CO₂. *Proceedings of the National Academy of Sciences of the United States of America (PNAS)*, **104**, 14014–14019.
- Gerten D, Lucht W, Schaphoff S, Cramer W, Hickler T, Wagner W (2005) Hydrologic resilience of the terrestrial biosphere. *Geophysical Research Letters*, **32**, L21408, doi: 10.1029/2005GL024247.
- Gerten D, Schabhoff S, Haberlandt U, Lucht W, Sitch S (2004) Terrestrial vegetation and water balance – hydrological evaluation of a dynamic global vegetation model. *Journal of Hydrology*, 286, 249–270.
- Gielen B, Calfapietra C, Lukac M *et al.* (2005) Net carbon storage in a poplar plantation (POPFACE) after three years of free-air CO₂ enrichment. *Tree Physiology*, **25**, 1399–1408.
- Haxeltine A, Prentice IC (1996a) A general model for light-use efficiency of primary production. *Functional Ecology*, **10**, 551–561.
- Haxeltine A, Prentice IC (1996b) BIOME3: an equilibrium terrestrial biosphere model based on ecophysiological constraints, resource availability, and competition among plant functional types. *Global Biochemical Cycles*, **10**, 693–709.
- Hendrey GR, Ellsworth DS, Lewin KF, Nagy J (1999) A free-air enrichment system for exposing tall forest vegetation to elevated atmospheric CO₂. *Global Change Biology*, 5, 293–309.
- Hickler T, Eklundh L, Seaquist JW et al. (2005) Precipitation controls Sahel greening trend. *Geophysical Research Letters*, 32, L21415, doi: 10.1029/2005GL024370.
- Hickler T, Prentice IC, Smith B, Sykes MT, Zaehle S (2006) Implementing plant hydraulic architecture within the LPJ dynamic global vegetation model. *Global Ecology & Biogeography*, **15**, 567–577.
- Hickler T, Smith B, Sykes MT, Davis M, Sugita S, Walker K (2004) Using a generalized vegetation model to simulate vegetation dynamics in northeastern USA. *Ecology*, **85**, 519–530.
- Hungate BA, Dukes JS, Shaw MR, Luo Y, Field CB (2003) Nitrogen and climate change. *Science*, **302**, 1512–1513.
- Jordan DB, Ogren WL (1984) The CO₂/O₂ specifity of ribulose-1,5-biphosphate concentration, pH and temperature. *Planta*, **161**, 308–313.
- King JS, Kubiske ME, Pregitzer KS, Hendrey GR, McDonald EP, Giardina CP, Quinn VS (2005) Tropospheric O₃ compromises net primary production in young stands of trembling aspen, paper birch and sugar maple in response to elevated atmospheric CO₂. New Phytologist, **168**, 623–636.
- Kirschbaum MUF (1994) The sensitivity of C₃ photosynthesis to increasing CO₂ concentration: a theoretical analysis of its dependence on temperature and background CO₂ concentration. *Plant, Cell and Environment*, **17**, 747–754.
- Körner C (2000) Biosphere responses to CO₂ enrichment. *Ecological Applications*, **10**, 1590–1619.
- Körner C, Asshoff R, Bignucolo O *et al.* (2005) Carbon flux and growth in mature deciduous forest trees exposed to elevated CO₂. *Science*, **309**, 1360–1362.
- Körner C, Diemer M, Schäppi B, Zimmermann L (1996) Response of alpine vegetation to elevated CO₂. In: *Carbon Dioxide*

© 2008 The Authors

Journal compilation © 2008 Blackwell Publishing Ltd, Global Change Biology, 14, 1531–1542

and Terrestrial Ecosystems (eds Koch GW, Mooney HA), pp. 177–196. Academic Press, San Diego.

- Liberloo M, Calfapietra C, Lukac M *et al.* (2006) Woody biomass production during the second rotation of a bio-energy *Populus* plantation increases in a future high CO₂ world. *Global Change Biology*, **12**, 1094–1106.
- Long SP (1991) Modification of the response of photosynthetic productivity to rising temperature by atmospheric CO₂ concentrations: has its importance been underestimated? *Plant, Cell and Environment*, **14**, 729–739.
- Lucht W, Prentice IC, Myneni RB *et al.* (2002) Climatic control of the high-latitude vegetation greening trend and Pinatubo effect. *Science*, **296**, 1687–1689.
- McCarthy HR, Oren R, Finzi AC, Johnsen KH (2006) Canopy leaf area constrains CO₂-induced enhancement of productivity and partitioning among aboveground carbon pools. *Proceedings of the National Academy of Sciences of the United States of America (PNAS)*, **103**, 19356–19361.
- McGuire AD, Sitch S, Clein JS *et al.* (2001) Carbon balance of the terrestrial biosphere in the twentieth century: analyses of CO₂, climate and land use effects with four process-based ecosystem models. *Global Biogeochemical Cycles*, **15**, 183–206.
- Melillo JM, McGuire AD, Kicklighter DW, Moore B, Vorosmarty CJ, Schloss AL (1993) Global climate change and terrestrial net primary production. *Nature*, 363, 234–240.
- Mitchell TD, Jones PD (2005) An improved method of constructing a database of monthly climate observations and associated high-resolution grids. *International Journal of Climatology*, **25**, 693–712.
- Moore DJP, Aref S, Ho RM, Pippen JS, Hamilton JG, De Lucia EH (2006) Annual basal area increment and growth duration of *Pinus taeda* in response to eight years of free-air carbon dioxide enrichment. *Global Change Biology*, **12**, 1367–1377.
- Morales P, Hickler T, Smith B, Sykes MT, Rowell DP (2007) Changes in European ecosystem productivity and carbon stocks driven by regional climate model outputs. *Global Change Biology*, **13**, 108–122.
- Morison JIL, Lawlor DW (1999) Interactions between increasing CO₂ concentration and temperature on plant growth. *Plant, Cell and Environment*, **22**, 659–682.
- Norby RJ, DeLucia EH, Gielen B *et al.* (2005) Forest response to elevated CO₂ is conserved across a broad range of productivity. *Proceedings of the National Academy of Sciences of the United States of America*, **102**, 18052–18056.
- Norby RJ, Hanson PJ, O'Neil EG *et al.* (2002) Net primary productivity of a CO₂ enriched deciduous forest and the implications for carbon storage. *Ecological Applications*, **12**, 1261–1266.
- Norby RJ, Iversen CM (2006) Nitrogen uptake, distribution, turnover, and efficiency of use in a CO₂-enriched Sweetgum forest. *Ecology*, **87**, 5–14.
- Norby RJ, Luo Y (2004) Evaluating ecosystem responses to rising atmospheric CO₂ and global warming in a multi-factor world. *New Phytologist*, **162**, 281–293.
- Norby RJ, Sholtis JD, Gunderson CA, Jawdy SS (2003) Leaf dynamics of deciduous forest canopy: no response to elevated CO₂. *Oecologia*, **136**, 574–584.
- Norby RJ, Wullschleger SD, Gunderson CA, Johnson DW, Ceulemans R (1999) Tree responses to rising CO₂ in field experi-

ments: implications for the future forest. *Plant, Cell and Environment*, **22**, 683–714.

- Nowak RS, Ellsworth DS, Smith SD (2004) Functional responses of plants to elevated atmospheric CO₂: do photosynthetic and productivity data from FACE experiments support early predictions? *New Phytologist*, **162**, 253–280.
- Oren R, Ellsworth DS, Johnsen KH *et al.* (2001) Soil fertility limits carbon sequestration by forest ecosystems in a CO₂-enriched atmosphere. *Nature*, **411**, 469–472.
- Pan Y, Mellilo JM, McGuire AD *et al.* (1998) Modeled responses of terrestrial ecosystems to elevated atmospheric CO₂: a comparison of simulations by biogeochemistry models of the Vegetation/Ecosystem Modeling and Analysis Project (VE-MAP). *Oecologia*, **114**, 389–404.
- Prentice IC, Farquhar GD, Fasham MJR et al. (2001) The carbon cycle and atmospheric carbon dioxide. In: Climate Change 2001: The Scientific Basis. Contribution of the Working Group I to the Third Assessment Report of the International Panel of Climate Change (eds Houghton JT, Ding Y, Griggs DJ et al.), pp. 183–239. Cambridge University Press, Cambridge.
- Prentice IC, Sykes MT, Cramer W (1993) A simulation model for the transient effects of climate change on forest landscapes. *Ecological Modelling*, **65**, 51–70.
- Rustad L, Campbell J, Marion G *et al.* (2001) A meta-analysis of the response of soil respiration, net nitrogen mineralization, and aboveground plant growth to experimental ecosystem warming. *Oecologia*, **126**, 543–562.
- Schäfer KVR, Oren R, Ellsworth DS *et al.* (2003) Exposure to enriched CO₂ atmosphere alters carbon assimilation and allocation in a pine forest ecosystem. *Global Change Biology*, **9**, 1378–1400.
- Sitch S, Smith B, Prentice IC *et al.* (2003) Evaluation of ecosystem dynamics, plant geography and terrestrial carbon cycling in the LPJ dynamic global vegetation model. *Global Change Biology*, **9**, 161–185.
- Smith B, Prentice IC, Sykes MT (2001) Representation of vegetation dynamics in the modelling of terrestrial ecosystems: comparing two contrasting approaches within European climate space. *Global Ecology & Biogeography*, **10**, 621–637.
- Thonicke K, Vevevsky S, Sitch S, Cramer W (2001) The role of fire disturbance for global vegetation dynamics: coupling fire into a dynamic global vegetation model. *Global Ecology & Biogeography*, **6**, 483–495.
- Tissue DT, Oechel WC (1987) Response of *Eriophorum vaginatum* to elevated CO_2 and temperature in the Alaskan tussock tundra. *Ecology*, **68**, 401–410.
- Wullschleger SD, Tschaplinski TJ, Norby RJ (2002) Plant water relations at elevated CO₂ – implications for water-limited environments. *Plant, Cell and Environment*, **25**, 319–331.
- Zaehle S, Sitch S, Smith B, Hatterman F (2005) Effects of parameter uncertainties on the modelling of terrestrial biosphere dynamics. *Global Biogeochemical Cycles*, **19**, GB3020, doi: 10.1029/2004GB002395.
- Zotz G, Pepin S, Körner C (2005) No down-regulation of leaf photosynthesis in mature forest trees after three years of exposure to elevated CO₂. *Plant Biology*, **7**, 369–374.

| Appen | dix | A |
|-------|-----|---|
|-------|-----|---|

| | ORNL | FACTS I | FACTS II | PopFACE |
|--|--|---|---|--|
| Location | Oak Ridge, Tennessee, USA | Durham, North Carolina, USA | Rhinelander, Wisconsin, USA | Toscana (Viterbo), Italy |
| Latitude, longitude | 35°54′N, 84°20′W | 35°58′N, 79°05′W | 45°40′N, 89°37′N | 42°22'N, 11°48'E |
| Observed LAI* | 5.1 | 3.4 | 2.7-3.4 | 4.6-7.4 |
| Modeled LAI [†] | 4.4 | 2.8 | 3.4 | 7.7 |
| Observed biomass $(g C m^{-2})$ | 5921‡ | 5394§ | ~ 800 †¶ | $749 \parallel$ |
| Modeled biomass $(g C m^{-2})$ | 2080‡ | 4200§ | 1049† | 2164 |
| Observed NPP $(g C m^{-2} yr^{-1})^{+**}$ | 994 (86) | 891 (252) | 330 (70) | 1451 (257††) |
| Modeled NPP $(g C m^{-2} yr^{-1})^{+**}$ | 758 (94) | 778 (15) | 671 (96) | 1125 (381) |
| Main species | Liquidambar styraciflua L. | Pinus taeda L. | Populus tremuloides Michx. | Populus alba L., P. nigra L., P. × euramericana Dode Gunier |
| Model PFT | Temperate, intermediate shade-tolerant, broad- leaved tree | Temperate, intermediate shade-tolerant, needle- leaved tree | Boreal, shade- intolerant, broad- leaved tree | Temperate, shade-intolerant, broad-leaved tree |
| Plantation year | 1988 | 1983 | 1997 | 1999 |
| Sapling density (per hectare) | 3600 | 1700 | 10 000 | 10 000 |
| Elevated CO ₂ concentration (ppmv) | 550 | 580 | 580 | 550 |

| Table A1 | Modeled | and | observed | FACE | site | characteristics |
|-----------|---------|-----|-----------|---------|------|-----------------|
| Iuvic III | moucicu | unu | UDDCI VCU | IIICL . | onc. | cianacteristics |

*LAI, leaf area index; data from Norby et al. (2005).

[†]Value for ambient CO₂, averaged for all years used in the data-model intercomparison (Fig. 1).

[‡]Value for ambient CO₂ for the year 1998 (http://public.ornl.gov/face/synthesis/N_Synthesis-CDIAC.htm).

§Value for ambient CO₂ for the year 1998, excluding hardwoods (Hamilton *et al.*, 2002).

¶King et al. (2005; Fig. 1).

||Value for 2001, averaged for three Populus species (Calfapietra et al., 2003).

**Values only for ambient CO₂, confidence intervals (treating each year as one measurement) given in parentheses.

 \dagger Data from the three *Populus* species and 2 years treated as independent data points (N = 6).

| Table A2 | Set of plant function | al types (PFTs) i | n the LPJ-DGVM | (Sitch et al., 2 | 2003) was e | nhanced by | subdividing e | ach Pl | FT into |
|-------------|------------------------|---------------------|-----------------------|------------------|-------------|--------------|---------------|--------|---------|
| shade-tole: | rance classes (Smith a | t al., 2001; Hickle | er et al., 2004), res | ulting into the | e following | list of PFTs | | | |

| PFT | Examples of typical taxa | |
|---|--------------------------|--|
| Tropical, shade-tolerant, broad-leaved, evergreen tree | | |
| Tropical, shade-intolerant, broad-leaved, evergreen tree | | |
| Tropical, intermediate shade-tolerant, broad-leaved, rain-green tree | | |
| Temperate, shade-tolerant, broad-leaved, deciduous tree | Fagus | |
| Temperate, intermediate shade-tolerant, broad-leaved, deciduous tree | Quercus, Tilia, Fraxinus | |
| Temperate, shade-intolerant, broad-leaved, deciduous tree | Betula, Populus | |
| Temperate, intermediate shade-tolerant, needle-leaved, evergreen tree | Pinus | |
| Temperate, intermediate shade-tolerant, broad-leaved, evergreen tree | Quercus | |
| Boreal, shade-tolerant, needle-leaved, evergreen tree | Picea | |

1540 T. HICKLER *et al.*

Table A2. (Contd.)

| PFT | Examples of typical taxa |
|--|--------------------------|
| Boreal, intermediate shade-tolerant, needle-leaved, evergreen tree | Pinus, Abies |
| Boreal, shade-intolerant, broad-leaved, deciduous tree | Betula, Populus |
| Boreal, shade-intolerant, needle-leaved, deciduous tree | Larix |
| Herbaceous with C_3 photosynthesis | |
| Herbaceous with C ₄ photosynthesis | |

Table A3 PFTs (Table A2) were distinguished by the following parameters (Smith et al., 2001)

| Leaf morphology | Needle-leaved | Broad-leaved | | |
|---|---------------|--------------|------------|--|
| Minimum canopy conductance (mm s ⁻¹) | 0.3 | 0.3 | | |
| Leaf turnover rate (year ⁻¹) | 0.5* | | 1.0 | |
| Fine root turnover rate (year ⁻¹)† | 0.7 | | 0.7 | |
| Leaf to sapwood cross-sectional area ratio‡ | 3000 (1500§) | | 4000 | |
| $k_{\rm allom1}$ ¶ | 150 | 200 | | |
| Climatic range | Boreal | Temperate | Tropical | |
| Optimal temperature range for photosynthesis (°C) | 10–25 | 15–25 | 20–30 | |
| Tissue respiration rate at 10 °C (g C g N ⁻¹ day ⁻¹) | 0.122 | 0.066 | 0.011 | |
| Shade tolerance | Tolerant | Intermediate | Intolerant | |
| Sapwood to heartwood conversion rate (year ⁻¹) | 0.03 | 0.05 | 0.08 | |
| Growth efficiency mortality parameter $(g C m^{-2} yr^{-1})^{**}$ | 90 | 100 | 110 | |
| Maximum establishment rate (saplings $yr^{-1}m^{-2}$)†† | 0.05 | 0.1 | 0.15 | |
| Minimum PAR at forest floor for establishment $(MJ m^{-2} day^{-1})$ | 1.0 | 1.5 | 2.0 | |
| Recruitment shape parameter ^{‡‡} | 0.8 | 1.3 | 2.0 | |

*With the exception of the boreal shade-intolerant needle-leaved tree (Table A2), which shed its leaves every year and therefore was assigned a leaf turnover rate of 0.5.

†In the original model version (Smith *et al.*, 2001), the root turnover rate was coupled with the leaf turnover rate. Here, one value was chosen for all PFTs, based on Vogt *et al.* (1996) and Li *et al.* (2003).

[‡]The original value of 8000 was replaced by lower values, following Martin *et al.* (1998), Becker *et al.* (1999), Eamus & Prior (2001), Tyree & Ewers (1991), McDowell *et al.* (2002).

§At the FACTS I site, a value of 1500 was used for *Pinus taeda*. This value is slightly higher than the values measured at the Duke forest by Naidu *et al.* (1998), which were approximately 1000, but using 1500 improved the modeled relationship between LAI and aboveground biomass.

 $\P k_{\text{allom1}}$ is an allometric parameter, determining the relationship between stem diameter and crown area [Sitch *et al.*, 2003; Eqn (4)]; a lower value implies smaller crowns for a given diameter.

||In LPJ-GUESS, different shade-tolerance classes are related to a number of parameters, which define a strategic trade-off: trees can either allocate to fast growth and high maximum recruitment; or to functions that permit survival at low growth rates resulting from intense shading by neighbors (Hickler *et al.*, 2004).

**Growth-efficiency-related mortality was modeled as an s-shaped function of the net primary productivity (NPP) of a tree cohort divided by its leaf area, averaged over the last 5 years (Prentice *et al.*, 1993). A growth efficiency mortality parameter (param_mort_{greff}) denotes the turning point of the curve defined by

 $mort_{greff} = 0.1/(1 + (greff_{5 vears}/param_mort_{greff})^5),$

where greff_mort represents the fraction of the cohort that will die, and $greff_{5 years}$ represents the growth efficiency over the last 5 years.

 \dagger Relative values, total maximum background establishment of all PFTs together was set to 0.3 saplings yr⁻¹ m⁻². Note that, for modeling of the FACE sites, establishment was prescribed (see 'Material and methods').

‡‡High values indicating strongly reduced establishment, as growth conditions at the forest floor become unfavorable as a result of low PAR levels (Fulton, 1991).



Fig. A1 Modeled relationship between NPP enhancement (Δ NPP) and environmental factors: (a) mean annual temperature; (b) mean annual actual/potential evapotranspiration (AET/PET). Values are averaged over 1996–2002.



Fig. A2 Effects of elevated CO₂ and temperature on lightsaturated gross daily photosynthesis (A_{sat}) of a temperate tree, as predicted by the photosynthesis module of LPJ-GUESS. The standard model is shown in black, and gray denotes the same simulations with k_c (Michaelis constant for Rubisco for CO₂), k_o (Michaelis constant of Rubisco for O₂), and τ (CO₂/O₂ specificity ratio) held constant at 25 °C values (i.e. no physiological temperature dependence; see Haxeltine & Prentice, 1996a for parameter values): (a) absolute values and (b) enhancement of A_{sat} in % if CO₂ is increased from 360 to 560 ppmv. At low temperatures and above a PFT-specific maximum temperature, photosynthesis is also inhibited by an empirical PFT-specific temperature-inhibition function (Sitch *et al.*, 2003). Day length was set to 12 h, and the photosynthetically active radiation (PAR) to 1500 µmol m⁻² s⁻¹.

References

- Becker P, Tyree MT, Tsuda M (1999) Hydraulic conductance of angiosperms versus conifers: similar transport sufficiency at the whole-plant level. *Tree Physiology*, **19**, 445–452.
- Eamus D, Prior L (2001) Ecophysiology of trees of seasonally dry tropics: comparisons among phenologies. *Advances in Ecological Research*, **32**, 113–197.
- Fulton MR (1991) Adult recruitment as a function of juvenile growth rate in size-structured plant populations. *Oikos*, **62**, 102–105.
- Hamilton JG, DeLucia ED, George K, Naidu SL, Finzi AC, Schlesinger WH (2002) Forest carbon balance under elevated CO₂. *Oecologia*, **131**, 250–260.
- Li Z, Kurz WA, Apps MJ, Beukema SJ (2003) Belowground biomass dynamics in the Carbon Budget Model of the Canadian Forest Sector: recent improvements and implications for the estimation of NPP and NEP. *Canadian Journal of Forest Research*, 33, 106–126.
- Martin JG, Kloeppel BD, Schaefer TL, Kimbler DL, McNulty SG (1998) Aboveground biomass and nitrogen allocation of ten deciduous southern Appalachian tree species. *Canadian Journal* of Forest Research, 28, 1648–1659.
- McDowell N, Barnard H, Bond B *et al.* (2002) The relationship between tree height and leaf area: sapwood area ratio. *Oecologia*, **132**, 12–20.
- Naidu SL, DeLucia EH, Thomas RB (1998) Contrasting patterns of biomass allocation in dominant and suppressed loblolly pine. *Canadian Journal of Forest Research*, **28**, 1116–1124.
- Tyree MT, Ewers FW (1991) The hydraulic architecture of trees and other woody plants. *New Phytologist*, **119**, 345–360.
- Vogt KA, Vogt DJ, Palmiotto PA, Boon P, O'Hara J, Asbjornson H (1996) Review of root dynamics in forest ecosystems grouped by climate, climatic forest type and species. *Plant and Soil*, 187, 159–219.